

## Preface: conservation of european ponds-current knowledge and future needs

Maria R. Miracle<sup>1,\*</sup>, Beat Oertli<sup>2</sup>, Régis Céréghino<sup>3</sup> & Andrew Hull<sup>4</sup>

<sup>1</sup> Dept. Microbiologia i Ecologia. Institut Cavanilles de Biodiversitat i Biologia Evolutiva. University of Valencia. 46100-Burjassot (Valencia). Spain.

<sup>2</sup> hepia Geneva University of Applied Sciences Western Switzerland, CH-1254 Jussy-Geneva, Switzerland.

<sup>3</sup> Université de Toulouse, EcoLab Laboratoire d'Ecologie Fonctionnelle, UMR5245, 118 route de Narbonne, F-31062 Toulouse cedex 9, France.

<sup>4</sup> Liverpool John Moores University. Clarence St. L3 5UG, Liverpool, UK.

\* Corresponding author: rosa.miracle@uv.es

### ABSTRACT

#### Preface: conservation of european ponds-current knowledge and future needs

Ponds are common elements of the landscape with an important role in the global processes of biosphere and biodiversity preservation. Recent research indicates that ecological characteristics of ponds are different from other inland water systems, but scientific knowledge is still insufficient and poor compared to lakes and rivers. Therefore, whilst indicators and conservation tools have been developed for most aquatic systems, there is also a gap between existing basic information on pond ecology and applied research. The European Pond Conservation Network (EPCN) with the aim of strengthening the links between basic and applied research and pond management organized its 3<sup>rd</sup> biennial meeting in Valencia (Spain) with the theme "Pond conservation: from science to practice". We present a selection of papers from this conference, which cover the three main topics of the sessions: (1) Management and conservation in practice, (2) Pond ecology at different scales and (3) Temporary ponds. The articles presented develop techniques for assessing the ecological status of this type of ecosystems, evidence the importance of ponds in a global scale, indicate that their conservation must take into account their spatial arrangement in networks, discuss environmental factors that are relevant to biodiversity conservation and provide information on different research areas such as biogeochemical processes, evolution of aquatic biota and community ecology.

**Key words:** Ponds, biodiversity, conservation, temporary ponds, global change.

### RESUMEN

#### Prefacio: conservación de las charcas europeas-conocimiento actual y necesidades futuras

*Las charcas son elementos habituales del paisaje que tienen un importante papel en los procesos globales de la biosfera y en la conservación de la biodiversidad. Investigaciones recientes indican que las características ecológicas de las charcas son diferentes de las de otros sistemas acuáticos, pero los conocimientos científicos de ellas son todavía insuficientes y pobres comparados con los de los lagos y ríos. Por lo tanto, mientras que hay un desarrollo avanzado de herramientas para la conservación de la mayoría de los ecosistemas acuáticos, subsiste un retraso entre los conocimientos básicos de ecología de las charcas y los aspectos aplicados para su correcta gestión. La Red Europea para la conservación de las charcas (EPCN) con el objetivo de estrechar la relación entre el conocimiento fundamental y aplicado y la gestión de las charcas organizó su tercera reunión bienal en Valencia (España) con el lema "Conservación de las charcas: de la ciencia a la práctica". Presentamos aquí una selección de los trabajos expuestos cubriendo los tres tópicos principales de las sesiones: (1) Gestión y conservación en la práctica, (2) ecología de las charcas a diferentes escalas y (3) charcas temporales. Los artículos presentados desarrollan técnicas para la identificación del estado ecológico de este tipo de ecosistemas, ponen de manifiesto la importancia de las charcas en los procesos globales, indican que para su conservación hay que considerar su distribución espacial en redes, discuten los factores ambientales relevantes para la conservación de la biodiversidad y proporcionan información sobre diferentes áreas de investigación como procesos biogeoquímicos, evolución de los organismos acuáticos y ecología de comunidades.*

**Palabras clave:** Charcas, biodiversidad, conservación, charcas temporales, cambio global.

## INTRODUCTION

In Europe, ponds are the most widespread aquatic habitat and collectively dominate the total area of continental standing waters. This, that is evident especially in Mediterranean countries where lakes are very scarce, has not been taken into account in local environmental studies and even less in studies of biosphere plumbing. The “emerging role” of ponds is in the title of the first contribution to this issue (Downing 2010), which makes evident, based on recent and improved data, that ponds collectively not only have more surface area than large lakes, but are also more important in storing carbon than large lakes, thus having a significant role in the Earth’s carbon balance and climate change. In addition, ponds also play many other valuable roles such as enhancing biodiversity, not only of aquatic organisms but also of terrestrial organisms that depend directly on these ecosystems as well as other indirect beneficial effects such as mitigating diffuse pollution or regulating temperature and humidity. In terms of regional diversity, a network of ponds has been found to make a greater contribution than lakes or rivers (Biggs *et al.*, 2005) and the existence of important interactions between species composition of different pond sites have been appreciated, when large spatial scales are considered (Briers & Biggs 2005). However, knowledge on ponds is only beginning and since recent studies have evidenced marked differences with lakes, we are aware that knowledge is insufficient and much lower when compared to other aquatic systems. There is a need therefore for further research on the organization and processes not only within ponds, but also among them.

Despite the recent increase in the interest of ponds and awareness of their vulnerability to degradation and fast disappearance of many of them, their protection is still inadequate. For instance, the most substantial piece of water legislation constituted to protect our waters, the European Water Framework Directive, does not apply to water bodies of less than 50 ha, in most Member States, although in some nations, as in Spain, a few smaller lakes have been exceptionally included, due to the fact of the low number

of natural water masses with an area greater than the 50 ha. Accordingly, this does not include any additional protection for important ponds. Some ponds, however, are protected under European Community legislation as providing a home for protected habitats listed in Annex I and protected species listed in Annex II and Annex IV of the Habitats Directive 92/43/EEC, mostly to the benefit of Mediterranean ponds. One step forward, however, would be to modify the Directive to recognize ponds or pond areas as an additional water-body type to be protected (EPCN, Pond Manifesto, 2008). Large-scale loss of these habitats, especially in the more arid Southern European countries, will be critical not only for conservation of aquatic and amphibious organisms but also to ameliorate climate change and also to maintain a connected landscape, because ponds, although small, constitute a series of vital stepping stones through the landscape as well as providing many benefits to surrounding ecosystems.

On the other hand, a number of pond conservation initiatives have been undertaken in some countries. In order to strengthen these, coordinate their activities and develop a framework of theoretical and practical knowledge for pond conservation, the European Pond Conservation Network (EPCN), was established at the first European Pond Workshop in 2004. This workshop was held in Geneva (Switzerland), devoted to “*Conservation and monitoring of pond Biodiversity*” with the objective of synthesizing recent basic and applied knowledge on the topic. One of the main outcomes from this initial meeting was the launching of the EPCN “to promote the awareness, understanding and conservation of ponds in a changing European landscape” (Oertli *et al.*, 2004; 2005a). The EPCN is a European network of people and institutions involved in basic and applied scientific research on pond conservation as well as a range of stakeholders involved directly or indirectly in any aspect of pond conservation. The second European Pond Workshop was devoted to “*Conservation of pond biodiversity in changing European landscape*” and was held in Toulouse (France) in 2006 where the main objectives were focused on understanding pond ecology, the added value of ponds and pond

management (Nicolet *et al.* 2007, Céréghino *et al.* 2008). The working sessions of this meeting were used to formulate the Pond Manifesto (EPCN, 2008), which had already been drafted at the first European Pond Workshop in 2004. The Manifesto sets out the case for the conservation of ponds, reveals the threats they face and outlines a strategy for their conservation in Europe, based on the knowledge and experience of researchers and practitioners. The Manifesto was delivered at the third EPCN conference in Valencia (Spain) and can be downloaded from the website of the EPCN ([www.europeanponds.org](http://www.europeanponds.org)).

Since the first workshop the network has been considerably active and has held biennial meetings. This special issue provides a sample of the papers presented at third meeting of the EPCN in Valencia (2008). Another selection of papers from this meeting has been published in a special issue of *Hydrobiologia* (Oertli *et al.*, 2009) and will also be collected in a volume from the series “Developments in Hydrobiology” (together with papers from the second European Pond workshop published in *Hydrobiologia* 597, 2008).

### **THIRD EUROPEAN POND WORKSHOP: SPECIAL ISSUE CONTENT**

The third European Pond workshop called for contributions on *Pond conservation: from science to practice*, with the aim of bringing together researchers, managers and practitioners to exchange information, concerns and views on common topics under different perspectives to strengthen knowledge on pond ecosystems. It was organized in Valencia under the auspices of EPCN by the Generalitat Valenciana (Conselleria de Medi Ambient) as an action included in the European Union Life-Nature project on “Restoration of priority habitats for amphibians”. A total of 123 communications were presented, 38 as oral presentations and 85 as posters (which can be downloaded from [http://campus.hesge.ch/epcn/posters\\_valencia08.asp](http://campus.hesge.ch/epcn/posters_valencia08.asp)). The meeting was structured around three topics: (1) Management and conservation in practice, (2) Pond ecology at different scales and

(3) Temporary ponds. In addition, two special working sessions were included in the conference programme. The first session focussed on *Pond management success stories* and, after the presentation of case studies where successful management had been carried out, was devoted to understanding how we measure “success” and what could be learnt from management failures. It was proposed that the EPCN website could store pond management stories whether successful or not. The second session –*Linking pond management to scientific knowledge*– was focussed on ways in which better links could be established between scientists and practitioners in order to coordinate fundamental and applied research and develop management practices on a scientific basis. The main issue discussed was ways in which the flow of information between management and research could be improved. This question is important for two reasons. Firstly, practitioners usually do not publish the results of their practices and are therefore not available to the scientific community and, secondly, there is very little applied research on pond management in scientific projects.

The papers selected for this issue cover the three main topics of the meeting. The study of ponds in a global scale is a new and very desirable perspective, which was the theme of the 1<sup>st</sup> keynote lecture of the meeting. In this lecture Downing (2010), based upon recent developments in data acquisition and mathematical approaches, clearly demonstrates the importance of ponds in global cycles, since they are small but numerous with a disproportionately high intensity of many processes. This review paper updates and illustrates with numbers the global balance of burial and evasion of carbon and the role of ponds in carbon processing. It also opens a great array of suggestions on global limnology and ecology and shows the need to integrate ponds in any study of global processes in the biosphere. Ponds are important beyond their local and regional scale, playing a significant role in global biogeochemical cycles and biodiversity maintenance.

The growing interest in temporary environments was reflected in the 2<sup>nd</sup> keynote lecture in which Brendonck *et al.* (2010) started the session

on temporary ponds with a well documented review of a large series of studies that these authors had undertaken in a series of small ephemeral freshwater rock pools. They indicate how these pools, which usually occur in clusters with different spatial patterns, can be used as model systems to study biological, evolutionary and ecological processes. In addition to the valuable results from their studies together with methodological descriptions, their paper includes attractive conceptual approaches and perspectives on patterns of species dispersal, meta-populations and meta-communities, as well as disturbance and community succession. Recent work, based in part on metapopulation concepts (Hanski 1999) has evidenced the importance of the interactions, mainly through dispersion, between ponds forming part of networks (Briers & Biggs 2005). The heterogeneity and gradients of environmental characteristics that display many diminutive idiosyncratic ponds, highly affected by surrounding local factors of their small catchment area, maintain a high regional biodiversity (Jeffries 1998), which may be richer than in other aquatic systems such as rivers, streams or ditches (Williams *et al.*, 2004). Several contributions in past workshops (Cayrou & Céréghino 2005; Jeffries 2005; De Bie *et al.*, 2008; Oertli *et al.*, 2008) have reinforced the idea that pond networks –pondscapes–, should be considered in any conservation strategy and the spatial and temporal scales should be broadened when developing management proposals. This large scale view is especially significant in temporary ponds (Pretus, 2009). The benefits of the pond landscape view for temporary pond conservation are illustrated in this issue by Diaz-Paniagua *et al.* (2009) integrating published and new data to describe the high species richness and wide community assembly variation among different ponds and years, dependent on fine gradients of hydrological and/or other factors found in the large numbers of temporary ponds of Doñana National Park (Southern Spain).

The study of temporary waters is far less developed than the study of other aquatic habitats and basic descriptions of these habitats is vital. Temporary ponds are fluctuant environments. Fernandez-Alaez & Fernandez-Alaez (2010) ex-

plore in temporary and permanent ponds, as well, but subject to marked seasonal fluctuations, the drastic changes of main ions and nutrients; firstly, after waterlessness in summer and then after refilling in autumn and spring. Long-standing natural temporary ponds, with a long history of a more or less predictable hydrological pattern, have evolved to start the annual wetting with a highly structured community of relict species not found in any other habitat type. Biodiversity, including active and diapause stages, shapes a stable community that becomes active by relatively predictable environmental pulses and that follows a repetitive process where succession trends can be tracked year after year. This view is well exemplified by the study of seasonal changes, focused mainly on crustaceans, in Sincarcas pond (East Spain) by Sahuquillo & Miracle (2010). This pond constitutes a true biodiversity hot spot, where communities with a high percentage of endangered relict or rare species are still thriving nowadays (with respect to crustaceans, all groups of large branchiopods and three coexisting diaptomids). There are not many ponds left with such a high diversity in Europe. The same study indicates that the deepening of a nearby pond has led to impoverishment and disappearance of temporary water specialists. Thus, it is highly recommended that conservation be directed towards maintaining ancient natural ponds as they are, with interventions limited to regulate those activities that could have impacts in its watershed and to remove human activities out of its basin, i.e. out of all the potentially flooding land, albeit it might not replenish to whole capacity all the years. Although this land could go for long dry periods, it should not be considered a waste land neither a land that needs restoration, but an integral part of the pond, in both its aquatic or terrestrial phase, containing a seed and egg bank of both phases.

Ecological assessment and monitoring is a major topic in conservation that has seldom been developed in ponds. As we have noticed above, they are not considered in the European Water Framework Directive by many Member States. However conservation of ponds is a recognized need (Pond Manifesto) due to increasing impacts of environmental alterations as a result, for exam-

ple, of land use in a changing climate. The papers by Angelibert *et al.* (2010) and Indermuehle *et al.* (2010) constitute an advanced step in developing a tool based on a rigorous scientific framework but useful for the “on the ground” practitioners. They propose the IBEM index a simplification of the PLOCH assessment method (Oertli *et al.*, 2005b), which follows the methodology adopted by the European Water Framework Directive, thus the ratio to a reference state is translated into one of five quality classes. To facilitate the method of implementation, a website (<http://campus.hesge.ch/ibem>) enables the calculation of the index online, and provides support to users on both sampling and assessment methodologies. The IBEM-Index is a rapid assessment standardized method which gives an overall value of pond biodiversity and has proven to be successful in regional screenings or site monitoring in Switzerland as a good indicator of ecological quality. Standardized sampling techniques are one of the key questions to obtain good comparative assessment data, but it is very important to select those that minimize the impact of sampling processes on the ecosystem. In this sense, it is remarkable the contribution of Scher *et al.* (2010) testing the invertebrate sampling efficiency and representativeness of different and resourceful artificial substrates. In addition to that, the work highlights the importance of the artificial substrate type on its colonization by macroinvertebrates in lentic systems.

Ecological restoration is also one of the management measures; Anton & Armengol (2010) studied different restored ponds in Albufera Natural Park (Mediterranean Spain coastal area) in relation to zooplankton diversity. One of the conclusions is that the lapsed time since a pond is restored is an important factor for species composition and diversity; but seems to be an important factor mainly in the temporary systems, since the permanent ponds are less influenced.

One of the more drastic restoring measures is directed to the creation of new ponds and several works have indicated the success of this practice (Williams *et al.*, 2008). In this issue, Garmendía & Pedrola (2010) present a short applied

paper addressed to practitioners describing a simple water balance model and its application to a hypothetical wetland pond albeit forced with real meteorological data in an arid country. The model explores how pond depth and shape are important for determining pond hydroperiod. The creation of ponds or modification of natural ones has been an ancient practice to hold water for different uses mainly irrigation and cattle watering. It has been shown that artificial, more or less intensively used ponds, may sustain biodiversity at a regional scale in an agricultural landscape (Céréghino *et al.*, 2008), this being true even in highway stormwater detection ponds (Scher *et al.*, 2004). Wide farm pond landscapes can be found in many agricultural areas of dry countries. In this issue, Leon *et al.* (2010) based on a comparison of a large number of farm ponds in Andalucía (Southern Spain) with the protected natural wetlands of this region reinforced the same conclusions that farm ponds are important to preserve biodiversity in the agricultural landscape. Species richness and diversity in farm ponds with natural substrates reached similar levels than natural wetlands. However their results show very clearly that ponds constructed or rebuilt with artificial substrates (plastic or concrete) had significantly lower zooplankton species richness than ponds with a natural substrate.

Due to their small size, ponds are very sensitive to the surrounding landscape, and the landscape indicators (Gergel *et al.*, 2002) applied to stream ecology, such as percentage of agricultural land, could also be used to predict a variety of water chemistry parameter in ponds. In the present issue, there is also a contribution that highlights the influence of land uses in the catchment area, in the water chemistry and trophic level of ponds (Kuczyńska-Kippen & Joniak, 2010). Surrounding land use might as well have an effect on the size of planktonic organisms (Basinska *et al.*, 2010). The last mentioned paper, where the size of the rotifer *Filinia* is analyzed, shows that size not only varies according to land uses but also in relation to the type of habitat in the pond where they are found: open waters or among emergent or submerged aquatic vegetation.

## PERSPECTIVES

Interesting new lines of thought have been initiated in pond studies, in the first paper of this issue Downing (2010) argues convincingly that ponds are biogeochemically very active and taken collectively a large fraction of carbon sequestration resides in their sediments. However, much work is still needed to quantify carbon and nutrient cycling and storage to understand regional and global budgets of greenhouse gases, at multiple scales of space and time. Ponds are very common landscape elements which originate spatial heterogeneity and are subject to high temporal variability. McClain *et al.* (2003) defined biogeochemical hot spots and hot moments respectively as patches or episodes that show disproportionately high reaction rates relative to the surrounding matrix or longer intervening time periods and recognized that hot spot and hot moment activity is often enhanced at terrestrial-aquatic interfaces. Therefore pond networks are very important sites with these characteristics and their spatial arrangements must be considered in natural resources management. Over a quarter of a century ago, Likens (1984) indicated the importance to protect beyond the shore line, because inland waters are interconnected elements of the landscape (surface and subterranean waters, airshed, soils, aquatic and terrestrial organisms). Land use changes affect the hydrologic routing and associated processing of transported materials which may alter natural linkages and perturb pond ecology, thus conservation measures must use watershed-ecosystem approaches.

Hydrological variation and spatial arrangement of ponds is very important for aquatic and terrestrial biota as well; moreover spatial heterogeneity and pond connectivity may increase substantially species richness in a metacommunity structure. Also individual sites, despite their small size, have been recognized to be truly biodiversity hot spots. These ponds, probably remnants of past larger network systems, should be preserved as they are and conservation measures will have to be taken in the watershed if they are threatened by intensifying agriculture or other

land uses. Since we know that processes are logarithmic and hysteresis occurs in the response of aquatic ecosystems to external forcing (Sheffer, 1998) in many sites it may be urgent to prevent further irreversible alterations. In the case of eutrophication, a sudden shift may occur after long lasting pollution; when a threshold is exceeded the system is transmuted to an alternative state and it will not respond to decreased pollution loads, until loads are reduced considerably below the mentioned threshold. But then, the system response to cessation of pollution will not retrace the same trajectory to initial conditions and if losses of biodiversity occur associated to the point of injuring the seed and egg bank, it will never return to its original state. It is preferable to preserve natural sites than to have to recover degraded ecosystems later. Most ponds or pond areas have small catchment areas that facilitate the identification of impacts, so conservation approaches including catchment area could be easily incorporated. Recent projects, such as the identification of Important Areas for Ponds (IAP project), already started successfully in the UK ([www.pondconservation.org.uk/pond\\_hap/iap.htm](http://www.pondconservation.org.uk/pond_hap/iap.htm)) will fulfil the lack of information on these environments and encourage better protection at large scales of biodiversity and pond resources. Many ponds have been created or modified for farm use. There is now a challenge to think ecologically in the future construction or management of small artificial water bodies. In agreement to recent results, to preserve biodiversity, constructed ponds have to mimic natural systems. Among the more important factors to consider are the maintenance of natural substrates (Boavida 1999), hydrology, morphology and reduction of the contamination of inflowing waters.

In 2010, 'The Year of Biodiversity', the 4th EPCN Conference will be held in Berlin (Erkner), with the theme "Eyes of the Landscape-value of ponds in the 21<sup>st</sup> century". Its objective is to intensify exchange of experiences of pond experts from both, basic sciences and applied work on conservation and management to address the issues of the Pond Manifesto (2008), as the organizers indicated in their invitation to the Conference.

## ACKNOWLEDGEMENTS

We are very grateful to the organizers of the EPCN meeting in Valencia, especially to Ignacio Lacomba, Vicente Sancho and Benjamí Perez for the excellent organization of a very valuable meeting. We acknowledge the support of the Life-Nature project “Restoration of priority habitats for amphibians” (LIFE05/NAT/E/00060) and of the “Conselleria de Medi Ambient, Aigua, Urbanisme i Habitatge of the Generalitat Valenciana”. Thanks also to all the manuscript reviewers and to Joan Armengol (chief editor of *Limnetica*) and the “Asociación Iberica de Limnología” (AIL) for the publication of this special issue.

## REFERENCES

- ANGÉLIBERT, S., V. ROSSET, N. INDERMUEHLE & B. OERTLI. 2010. The pond biodiversity index “IBEM”: a new tool for the rapid assessment of biodiversity in ponds from Switzerland. Part I. Index development. *Limnetica*, 29: 93-104.
- ANTÓN-PARDO, M. & X. ARMENGOL. 2010. Zooplankton community from restored peridunal ponds in L’Albufera Natural Park. *Limnetica*, 29: 133-144.
- BASIŃSKA A., N. KUCZYŃSKA-KIPPEN, K. ŚWIDNICKI. 2010. The body size distribution of *Filinia longiseta* (Ehrenberg) in different types of small water bodies in the Wielkoposka region. *Limnetica*, 29: 171-182.
- BIGGS, J., P. WILLIAMS, P. WHITFIELD, P. NICOLET & A. WEATHERBY. 2005. 15 years of pond assessment in Britain: results and lessons learned from the work of Pond Conservation. *Aquatic Conserv.: Mar. Freshw. Ecosyst.*, 15: 693-714.
- BOAVIDA, M. J. 1999. Wetlands: most relevant structural and functional aspects. *Limnetica*, 17: 57-63.
- BRENDONCK, L., M. JOCQUE, A. HULSMANS & B. VANSCHOENWINKEL. 2010. Pools ‘on the rocks’: freshwater rock pools as model system in ecological and evolutionary research. *Limnetica*, 29: 25-40.
- BRIERS, R. A. & J. BIGGS. 2005. Spatial patterns in pond invertebrate communities. Separating environments and distance effects. *Aquatic Conserv.: Mar. Freshw. Ecosyst.*, 15: 549-557.
- CAYROU, J. & R. CÉRÉGHINO. 2005. Life-cycle phenology of some aquatic insects: implications for pond conservation. *Aquatic Conserv.: Mar. Freshw. Ecosyst.*, 15: 559-571.
- CÉRÉGHINO, R., A. RUGGIERO, P. MARTY, S. ANGÉLIBERT, 2008. Biodiversity and distribution patterns of freshwater invertebrates in farm ponds of a south-western French agricultural landscape. *Hydrobiologia*, 597: 43-51.
- CÉRÉGHINO, R., J. BIGGS, B. OERTLI & S. DECLERCK. 2008. The ecology of European ponds: defining the characteristics of a neglected freshwater habitat. *Hydrobiologia*, 597: 1-6.
- DE BIE, T., S. DECLERCK, K. MARTENS, L. DE MEESTER & L. BRENDONCK, 2008. A comparative analysis of cladoceran communities from different water body types: patterns in community composition and diversity. *Hydrobiologia*, 597: 19-27.
- DÍAZ-PANIAGUA, C., R. FERNÁNDEZ-ZAMUDIO, M. FLORENCIO, P. GARCÍA-MURILLO, C. GÓMEZ-RODRÍGUEZ, A. PORTHEAULT, L. SERRANO & P. SILJESTRÖM. 2010. Temporary ponds from Doñana National Park: A system of natural habitats for the preservation of aquatic flora and fauna. *Limnetica*, 29: 41-58.
- DOWNING, J. 2010. Emerging global role of small lakes and ponds. Little things mean a lot. *Limnetica*, 29: 9-24.
- EPCN 2008, Pond Manifesto, [www.europeanponds.org](http://www.europeanponds.org)
- FERNÁNDEZ-ALAEZ, C. & M. FERNÁNDEZ-ALAEZ. (2010) Temporary ponds of Eastern Spain: Limnological typology and human impact. *Limnetica*, 29: 59-74.
- GARMENDIA, A. & J. PEDROLA-MONFORT. 2010. Simulation model comparing the hydroperiod of temporary ponds with different shapes. *Limnetica*, 29: 145-152.
- GERGEL, SE., TURNER, M. G., MILLER J. R., STANLEY, E. H., MELACK, J. M. 2002. Landscape indicators of human impacts to riverine systems. *Aquatic Sci.*, 64: 118-128.
- HANSKI, I. 1999. Metapopulation Ecology. Oxford University Press. Oxford, UK. 313 pp.
- INDERMUEHLE, N., S. ANGÉLIBERT, V. ROSSET & B. OERTLI. 2010. The pond biodiversity index “IBEM”: a new tool for the rapid assessment of biodiversity in ponds from Switzerland. Part 2. Method description and examples of application. *Limnetica*, 29: 105-120.

- JEFFRIES, M. 1998. Pond macrophyte assemblages, biodiversity and spatial distribution of ponds in the Northumberland coastal plain, UK. *Aquatic Conserv.: Mar. Freshw. Ecosyst.*, 8: 657-667.
- JEFFRIES, M. 2005b. Small ponds and big landscapes: the challenge of invertebrate spatial and temporal dynamics for European pond conservation. *Aquatic Conserv.: Mar. Freshw. Ecosyst.*, 15: 541-547.
- KUCZYŃSKA-KIPPEN, N. & T. JONIAK. 2010. Chlorophyll *a* and physical-chemical features of small water bodies as indicators of land use in the Wielkopolska region (Western Poland). *Limnetica*, 29: 163-170.
- LEÓN, D., P. PEÑALVER, J. CASAS, M. JUAN, F. FUENTES, I. GALLEGO & J. TOJA. 2010. Zooplankton richness in farm ponds of Andalusia (Southern Spain). A comparison with natural wetlands. *Limnetica*, 29: 153-162.
- LIKENS, G. 1984. Beyond the shore line: a watershed-ecosystem approach. *Verh. Internat. Verein. Limnol.*, 22: 1-22.
- McCLAIN, M. E., E. W. BOYER, C. L. DENT, S. E. GERGEL, N. B. GRIMM, P. M. GROFFMAN, S. C. HART, J. W. HARVEY, C. A. JOHNSTON, E. MAYORGA, W. H. McDOWELL & G. PINAY. 2003. Biogeochemical Hot Spots and Hot Moments at the Interface of Terrestrial and Aquatic Ecosystems. *Ecosystems*, 6: 301-312.
- NICOLET P., A. RUGGIERO & J. BIGGS. 2007. Second European Pond Workshop: Conservation of pond biodiversity in a changing European landscape. *Ann. Limnol.-Int. J. Lim.*, 43: 77-80.
- OERTLI, B., D. AUDERSET JOYE, N. INDERMUEHLE, R. JUGE & J.-B. LACHAVANNE. 2004. 1st European Pond Workshop "Conservation and monitoring of pond biodiversity" *Arch. Sci.*, 57: 69-72.
- OERTLI, B., J. BIGGS, R. CÉRÉGHINO, P. GRILLAS, P. JOLY, & J.-B. LACHAVANNE. 2005 (a). Conservation and monitoring of pond biodiversity. *Aquatic Conserv.: Mar. Freshw. Ecosyst.*, 15: 535-540.
- OERTLI, B., D. AUDERSET JOYE, E. CASTELLA, R. JUGE, A. LEHMANN & J.-B. LACHAVANNE. 2005 (b). PLOCH: a standardized method for sampling and assessing the biodiversity in ponds. *Aquatic Conserv.: Mar. Freshw. Ecosyst.*, 15: 665-679.
- OERTLI, B., N. INDERMUEHLE, S. ANGÉLIBERT, H., HINDEN & A. STOLL. 2008. Macroinvertebrate assemblages in 25 high alpine ponds of the Swiss National Park (Cirque of Macun) and relation to environmental variables. *Hydrobiologia*, 597: 29-41.
- OERTLI, B., R. CÉRÉGHINO, A. HULL & M. R. MIRACLE. 2009. Pond conservation: from science to practice. *Hydrobiologia*, 634: 1-9.
- PRETUS, J. LL. 2009. Mediterranean temporary ponds: life histories for unwarranted offbeat environments. In: *International Conference on Mediterranean Temporary Ponds. Proceedings & Abstracts*. P. Fraga-Arguimbau (ed.). *Reserca*, 14: 23-35. Consell Insular de Menorca. Maó.
- SAHUQUILLO, M. & M. R. MIRACLE. 2010. Crustacean and rotifer seasonality in a Mediterranean temporary pond with high biodiversity (Lavajo de Abajo de Sinarcas, Eastern Spain). *Limnetica*, 29: 75-92.
- SCHER, O., P. CHAVAREN, M. DESPRAUX & A. THIÉRY, 2004. Highway stormwater detention ponds as biodiversity islands? *Archive des Sciences*, 57: 121-130.
- SCHER, O., K. E. MCNUTT & A. THIÉRY. 2010. Designing a standardised sampling method for invertebrate monitoring: a pilot experiment in a motorway retention pond. *Limnetica*, 29: 121-132.
- WILLIAMS, P., M. WHITFIELD, J. BIGGS, S. BRAY, G. FOX, P. NICOLET & D. SEAR, 2004. Comparative biodiversity of rivers, streams, ditches and ponds in an agricultural landscape in Southern England. *Biological Conservation*, 115: 329-341.
- WILLIAMS, P., M. WHITFIELD & J. BIGGS, 2008. How can make new ponds biodiverse? A case study monitored over 7 years. *Hydrobiologia*, 597: 137-148.

# Emerging global role of small lakes and ponds: little things mean a lot

John A. Downing\*

Ecology, Evolution & Organismal Biology, Iowa State University, Ames, IA, USA

\* Corresponding author: downing@iastate.edu

Received: 6/10/09

Accepted: 18/10/09

## ABSTRACT

### Emerging global role of small lakes and ponds: little things mean a lot

Until recently, small continental waters have been completely ignored in virtually all global processes and cycles. This has resulted from the neglect of these systems and processes by ecologists and the assumption that ecosystems with a small areal extent cannot play a major role in global processes. Recent inventories based on modern geographical and mathematical approaches have shown that continental waters occupy nearly twice as much area as was previously believed. Further, these inventories have shown that small lakes and ponds dominate the areal extent of continental waters, correcting a century-long misconception that large lakes are most important. The global importance of any ecosystem type in a process or cycle is the product of the areal extent and the intensity of the process in those ecosystems. Several analyses have shown the disproportionately great intensity of many processes in small aquatic ecosystems, indicating that they play an unexpectedly major role in global cycles. Assessments of the global carbon cycle underscore the need for aquatic scientists to view their work on a global scale in order to respond to the Earth's most pressing environmental problems.

**Key words:** Ponds, lakes, global limnology, carbon, lake size, sequestration.

## RESUMEN

### *La emergencia del papel global de los pequeños lagos y charcas: el gran significado de las pequeñas cosas*

*Hasta muy recientemente, las aguas continentales de pequeño volumen se han ignorado completamente en todos los procesos y ciclos globales. Esto ha sido el resultado de la poca consideración de estos ecosistemas y procesos por los ecólogos y de asumir que los ecosistemas que ocupan un área pequeña no juegan ningún papel importante en los procesos globales. Inventarios recientes basados en aproximaciones geográficas y matemáticas modernas indican que las aguas continentales ocupan casi el doble del área de lo que se creía anteriormente. Además, estos inventarios han mostrado que las charcas y lagunas de pequeñas dimensiones predominan en la extensión superficial de las aguas continentales, corrigiendo la concepción equivocada de todo un siglo de que los grandes lagos eran los más importantes. La importancia global de cualquier tipo de ecosistema en un proceso o ciclo es el producto de su superficie por la intensidad del proceso en el ecosistema. Diversos análisis han mostrado la intensidad desproporcionadamente grande de muchos procesos en los pequeños sistemas acuáticos, indicando su sorprendente papel primordial en los ciclos globales. Evaluaciones del ciclo global del carbono ponen de manifiesto la necesidad de que los ecólogos acuáticos tengan una visión de su trabajo a escala global, para poder responder a los problemas ambientales más preocupantes.*

**Palabras clave:** Charcas, lagos, limnología global, tamaño de los lagos, secuestro de carbono.

## INTRODUCTION

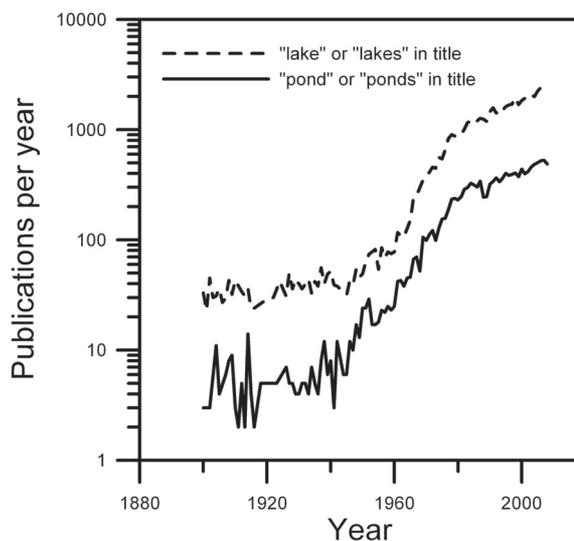
Ever since Halbfass (1914) and Thienemann's (1925) work cataloguing the lakes of the world,

science has assumed that the world's large lakes cover the most area and therefore are the most important to study (Downing *et al.* 2006, Downing & Duarte 2009). In spite of this long-standing

error of scientific reasoning (Downing 2009), our common, human experiences tell us that small things in life, society, or nature can be more important than their sizes imply. For example, part of the title of this article (“Little things mean a lot. . .”) comes from song lyrics by Edith Lindeman (no relation to Raymond) expressing that the tiny gestures people make have the most value. The 19<sup>th</sup> Swiss philosopher and poet, Henri-Frédéric Amiel, suggested that “What we call little things are merely the causes of great things” (Amiel 1893). Bruce Fairchild Barton, the American publicist, politician, and author wrote, “Sometimes when I consider what tremendous consequences come from little things... I am tempted to think there are no little things” (Barton 1917). The 18<sup>th</sup> century German scientist, satirist, and philosopher, Georg Christoph Lichtenberg, noted that “the tendency of people to consider small things as important has produced many great things” (Friederici 1978). We should not be misled by their small relative size into assuming that small lakes and ponds are unimportant. In *A Case of Identity* (Conan Doyle 1920), Sir Arthur Conan Doyle (speaking as Sherlock Holmes) suggested, “It has long been an axiom of mine that the little things are infinitely the most important.” Human experience suggests that we should expect the small parts of aquatic ecosystems, e.g., small lakes, ponds, puddles, marshes, and streams, to be of disproportionately great importance in world cycles and processes.

### Lakes, especially small ones, are ignored globally

Globally, lakes and ponds are generally ignored as being insignificant or are thought of only as reservoirs where water and materials are held for a short time before delivery to streams, rivers, and the oceans. Terrestrial ecologists, climatologists, and oceanographers tend to think of continental waters as “plumbing” that delivers or transports water, with little processing. Recently, this has been shown to be an incorrect assumption (Cole *et al.* 2007, Downing 2009, Tranvik *et al.* 2009). Further, scientists studying lentic waters have long known that they process glob-



**Figure 1.** Frequency analysis of use of “lake” or “lakes” versus “pond” or “ponds” in the title of scientific publications indexed by the Web of Science over the last century. Absolute frequency is dependent on the literature indexed by Web of Science and the completeness of index coverage. *Análisis de frecuencias de la utilización de la palabra “lake” o “lakes” versus “pond” o “ponds” en los títulos de las publicaciones científicas indexadas en la Web of Science durante el siglo pasado. Las frecuencias absolutas dependen de la bibliografía indexada y la cobertura de dicho índice.*

ally important materials. The concepts of nutrient and material retention and spiraling have been rudiments of limnology for several decades.

The study of small aquatic systems has lagged behind larger-lake limnology over much of the past century. An analysis of publications on “ponds” versus “lakes” in the publications indexed by *Web of Science* (Fig. 1) suggests the bias of ecologists and limnologists toward studying larger water bodies as well as the differential rates of growth of publications in these areas (see also Oertli *et al.* 2009). This analysis shows that studies titled as pond studies constitute only about 25% of the aquatic publications indexed in any given year. Further, although the rate of growth in the publication of pond studies increased at an average 19% per year from 1940–1980, lake studies increased extremely rapidly during the boom years of eutrophication remediation. Publications entitled as pond or lake studies have decelerated in the past decade, with rates of growth in pond analyses decelerating more than those of lakes.

**Table 1.** Analyses of global cycles and processes completely omitting any reference to ponds or small lakes. *Análisis de ciclos y procesos globales omitiendo totalmente cualquier referencia a charcas o lagunas.*

Cycle or budget	Reference
Carbon	(Goody & Walker 1972, Bolin 1983, Schimel <i>et al.</i> 1995, Intergovernmental Panel on Climate Change 2001, United States Climate Change Science Program 2003)
Energy/Radiation	(Christopherson 1994, Kiehl & Trenberth 1997, Hermann 2006)
Greenhouse gases	CO <sub>2</sub> : (Thorneloe <i>et al.</i> 2002) CH <sub>4</sub> : (Weissert 2000) N <sub>2</sub> O: (Seinfeld & Pandis 1998)
Nitrogen	(Rosswall 1983, Chameides & Perdue 1997, Bin-le <i>et al.</i> 2000, Roy <i>et al.</i> 2003, Raven <i>et al.</i> 2004)
Oxygen	(Cloud & Gibor 1970, Goody & Walker 1972, Walker 1980, Keeling <i>et al.</i> 1993)
Phosphorus	(Graham & Duce 1979, Richey 1983, Lerman 1988)
Silicon	(Goody & Walker 1972, Nelson <i>et al.</i> 1995, Tréguer <i>et al.</i> 1995)
Sulphur	(Freney <i>et al.</i> 1983, Raven <i>et al.</i> 2004)
Water	(Clarke 1991, Hinrichsen <i>et al.</i> 1998, Winter <i>et al.</i> 1998)

That small aquatic ecosystems are currently perceived as irrelevant to global problems is, however, undeniable. One needs only to look at schematic diagrams of various global material cycles to see that limnology and aquatic ecology have been left behind. Nowhere is this more obvious than in global analyses of the carbon cycle (e.g., Schimel *et al.* 1995). All continental waters are frequently absent from these global views. The carbon they store and any processing of this material they do (e.g., burial, emission) are completely omitted. Small, continental aquatic ecosystems are ignored in virtually all global views and processes (Table 1).

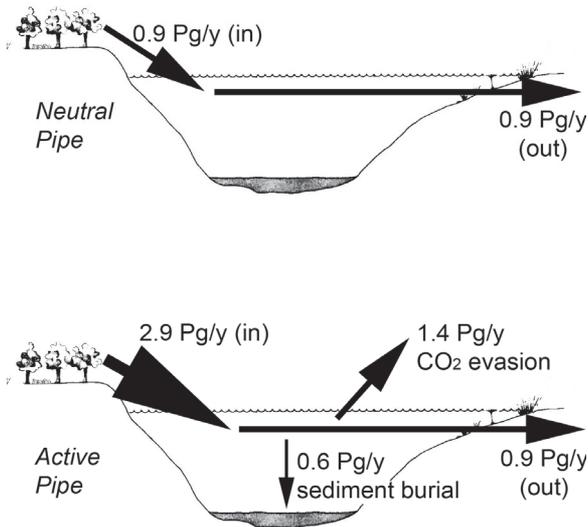
### Lakes, ponds, rivers, and streams are of global importance

Although they have been ignored, limnologists know that our systems are significant in global cycles. Nowhere is failing to consider them more serious than in the global carbon budget. Accuracy of estimation of the global carbon budget is critical because it will determine how effectively society can respond to the challenge of global climate change.

A few years ago, some of us attempted to integrate fragmentary knowledge on the role of inland waters into the global Carbon (C) cycle

(Downing *et al.* 2006, Cole *et al.* 2007). The information available at the time indicated that, far from being neutral conduits of C from lands to the sea, inland waters process large amounts of carbon buried in freshwater ecosystems or degassed to the atmosphere. Since that time, we have learned that the first calculations underestimated the area covered by virtually every category of inland waters (Downing *et al.* 2006, Downing 2009, Downing & Duarte 2009). Those estimates demonstrated that inland waters may process about 1 Pg/y (petagram/year) more C than was previously thought to be delivered to them. This was more than double the amount back-calculated as the landscape's contribution to rivers and the sea through the supposedly neutral conduit of inland waters. These numbers are being revised upward quite rapidly (e.g., Tranvik *et al.* 2009) and now show a very active processing of C by aquatic ecosystems (Fig. 2). Traditional analyses have calculated the loss of C from the landscape simply as the amount delivered to the sea by rivers but these calculations have ignored the role of inland waters in emitting and burying C.

Cole *et al.*'s (2007) calculations are being rapidly revised upward, underscoring the need for limnologists to engage in global limnology



**Figure 2.** Illustration of the quantitative and qualitative differences between the “neutral pipe” model suggesting the inland waters transport carbon without processing it, and the “active pipe” model (Cole *et al.* 2007) in which preliminary estimates of the global burial of C by aquatic ecosystems and the evasion of CO<sub>2</sub> by aquatic ecosystems is admitted. The original view of these models has been revised to reflect more recent data (Tranvik *et al.* 2009). This revision suggested that the large burial and evasion of carbon by aquatic ecosystems requires that export from land is almost three-times greater than previously believed. (Pg/y = 10<sup>15</sup> grams/year). *Esquema de las diferencias cuantitativas y cualitativas entre el modelo de “conducto neutro” en donde las aguas continentales transportan el carbono sin procesarlo y el modelo de “conducto activo” (Cole et al. 2007) en el cual se admite el entierro global de C y la liberación de CO<sub>2</sub> por los ecosistemas acuáticos. El esquema original de estos modelos se ha revisado para reflejar los datos más recientes (Tranvik et al. 2009). Esta revisión sugiere que el prominente entierro y liberación de C por los ecosistemas acuáticos, requiere que se exporte desde las zonas terrestres una cantidad casi tres veces mayor de lo que anteriormente se creía.*

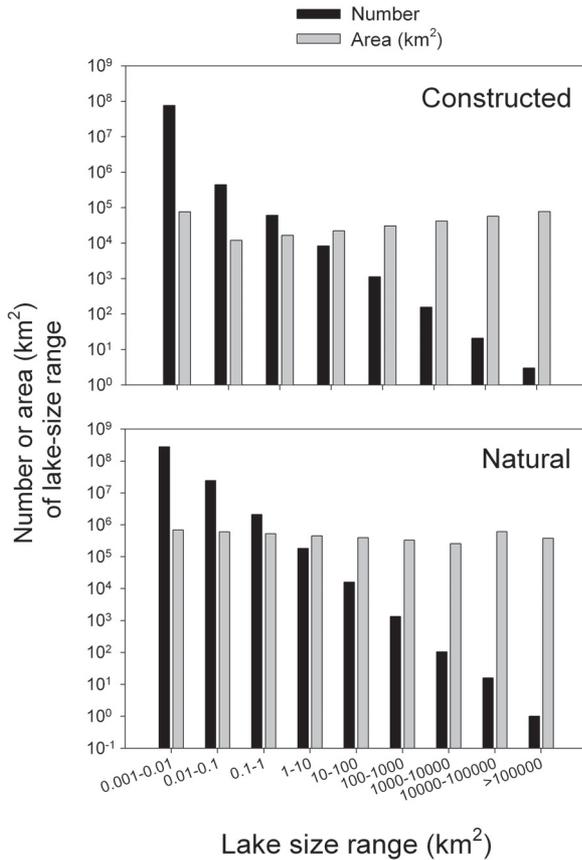
(Downing 2009). This lacuna is very obvious considering the under-emphasis of the global role of small aquatic ecosystems. The former view that Earth’s important compartments are ocean, atmosphere, and land, connected together by the assumed neutral pipes and conduits provided by large lakes and rivers was a major error. An accurate understanding of global cycles requires seeing the biosphere as a network of inter-connected metabolically active sites, including small lakes and ponds.

### Why might small lakes and ponds be very important?

It has recently been suggested that the global importance of any set of ecosystems is determined by the product of the amount of the biosphere they constitute and the intensity of the process of interest within them (Downing 2009). Downing (2009) also explored ways of “scaling-up” measurements made in small lakes and ponds for evaluating their global role. The global role of small lakes and ponds has been doubly missed in the past because the spatial extent of lakes has been underestimated as well as the fraction of the world’s lakes that are small (Lehner & Döll 2004, Downing *et al.* 2006).

An early inventory of the world’s lakes was first published in 1914 (Halbfass 1914) and was expanded to include August Thienemann’s analysis of the lakes of Europe (Thienemann 1925). At that time, Thienemann (1925) suggested that around 2.5 million km<sup>2</sup> or about 1.8 % of the land surface, is covered with lakes and ponds, and that global lake area is dominated by a few very large lakes (Downing 2009). This viewpoint was fundamentally unchanged for about 70 years (Schuiling 1977, Herdendorf 1984, Meybeck 1995, Kalff 2001) except that Robert Wetzel (1990) felt that the world’s lake area is dominated by small lakes and ponds (Downing *et al.* 2006).

Lehner and Döll (2004) performed a full inventory of world lakes by using GIS of satellite imagery to count all of the world’s moderately sized to large lakes, but could not count small lakes and ponds ( $\leq 0.1$  km<sup>2</sup>). Their data suggested a Pareto distribution (Pareto 1897, Vidondo *et al.* 1997) that appears to fit lake-size distributions down to 0.001 km<sup>2</sup> (Downing *et al.* 2006). A similar relationship was also found to fit the abundance and size-distribution of the world’s constructed lakes and analyses of regional data showed that constructed farm ponds bore a consistent relationship to agricultural land area and precipitation (Downing *et al.* 2006). These results suggest that there are 304 million natural lakes in the world and they cover about 4.2 million km<sup>2</sup>. This area is nearly twice that assumed by several others (Schlesinger 1997, Kalff 2001,



**Figure 3.** Global size distributions of numbers and land area covered by natural and constructed lakes. Data are re-plotted from the original publication (Downing *et al.* 2006). The figure shows that size distribution of natural lakes and constructed lakes are similar and that global lake area is dominated by small lakes, not large ones as 20<sup>th</sup> century analyses suggested (Halbfass 1914, Thienemann 1925, Schuiling 1977, Herdendorf 1984, Meybeck 1995). *Distribución global del número y superficie de los lagos construidos y naturales. Datos reproducidos de la publicación original (Downing et al. 2006). La figura muestra que el tamaño de los lagos naturales y los construidos es similar y que el área global está dominada por los lagos pequeños y no por los grandes como sugerían los análisis efectuados en el siglo XX.*

Wetzel 2001, Shiklomanov & Rodda 2003) and the area is more strongly dominated by small lakes and ponds (Fig. 3) than past analyses have suggested.

### How small are the smallest lakes and how long do they last?

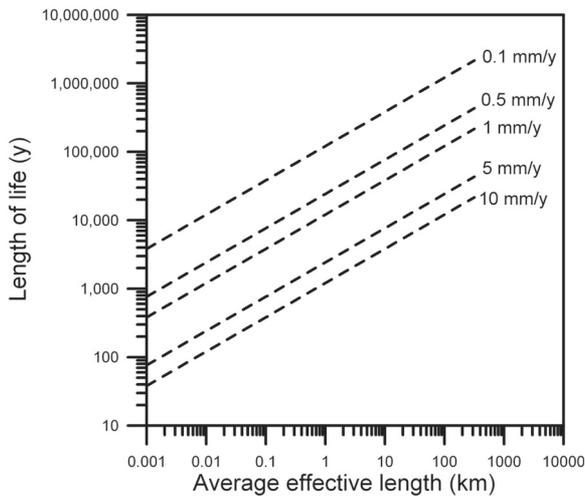
Many pond ecologists work on water bodies even smaller than the lowest interval on figure 3. If one

uses the Pareto distribution to project the number of water bodies on Earth in the range of 0.0001-0.001 km<sup>2</sup> (100-1000 m<sup>2</sup>), the result accentuates the dominance of small water bodies on continents. It is likely that there are about  $3.2 \times 10^9$  natural ponds in this size-range and they have an area of around 0.8 billion km<sup>2</sup>. Whether these ecosystems are permanently aquatic or become semi-terrestrial at certain times of the year, or whether they wax and wane over the course of geological time is not fully known. Our ability to catalog and map small features is, as yet, new, and we will learn how these small landscape features contribute to the interface of terrestrial and aquatic ecology.

Most of the Pareto distributions we have analyzed (Downing *et al.* 2006) had some curvature toward the small sizes of lakes, implying that they had been underestimated in inventories, removed from the landscape through erosion, deposition, and landscape alteration, or both. It seems quite likely that the residence time of small water bodies on a landscape may be low enough that some small systems disappear over time or are replaced by processes of pond formation. Some may be essentially hydric soils for part of the year. Any alteration of the land surface, including the filling of depressions can result in new small depressions that accumulate water and generate an aquatic ecosystem.

The intensive activity of small aquatic ecosystems and their dimensions make them more dynamic in time than large water bodies. I know, for example, of many small ponds that I knew as a child that are no longer part of the aquatic landscape. Likewise, however, I know of many modern small ponds that did not exist a few decades ago. One can estimate the relationship between the sizes of lakes or ponds and their likely life-spans following some assumptions about dimensions and morphometry. If the mean depth (m) of a lake is assumed to be  $12.1 \sqrt{L}$ , where  $L$  is the average of effective length and breadth (km) (Gorham 1958, Straškraba 1980), figure 4 shows the likely life-span of these lakes and ponds, assuming that lakes are elliptical in shape with length about double the breadth.

If sediment deposition is around 1 mm/y then very small lakes and ponds (< 0.01 km<sup>2</sup>) will



**Figure 4.** Potential life-time of aquatic ecosystems of a range of sizes. The calculations were based on assumed rates of sedimentation spanning the range of those observed in oligotrophic to eutrophic lakes and the assumptions that the mean depth (m) of a lake is around  $12.1 \sqrt{L'}$ , where  $L'$  is the average of effective length and breadth (km) (Gorham 1958, Straškraba 1980), and length is approximately double the breadth. *Duración potencial de los ecosistemas acuáticos de diferentes tamaños. Los cálculos se han basado en las diferentes tasas de sedimentación estimadas de las observadas en lagos, desde oligotróficos a eutróficos, y en el supuesto de que la profundidad media (m) de un lago sería  $12.1 \sqrt{L'}$ , en donde  $L'$  es la media de la longitud y anchura efectivas (Gorham 1958, Straškraba 1980), siendo la longitud aproximadamente el doble de la anchura.*

have lifetimes of <1000 y. In even more oligotrophic landscapes where sediment deposition rates are < 1 mm/y, small lakes and ponds might take 1000-10,000 y to disappear. In highly erodible, nutrient-enriched environments, however, substantially sized small lakes and ponds may disappear in a few decades through filling and succession. This temporal dynamic is a unique feature of the limnology of small lakes and ponds and accentuates our need to understand their function as well as their succession and origination.

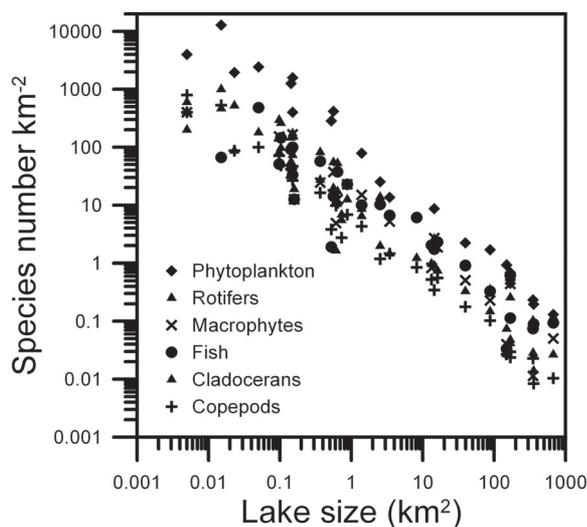
### Ponds and small lakes play an active global role

The global importance of any ecosystem type is determined by the product of the aerial extent of that ecosystem across the Earth and the intensity of processes in them, relative to other ecosystem types (Downing 2009). Indeed, the global

dominance of limnological processing only requires that these processes be more than 33-times greater (on an areal basis) in lakes than in terrestrial environments and more than 115-times greater than in the world's oceans. If globally important rates and processes are the same in small ( $\leq 1 \text{ km}^2$ ) lakes and ponds as they are in larger ones, small lakes and ponds constitute at least a third of the processing by aquatic ecosystems on the planet (Fig. 3). For small lakes and ponds to dominate inland aquatic processing, rates and processes in small systems need only be double those seen in larger ones. Knowledge of the “intensity” of processes is an important need in order to participate in global science.

Many aquatic rates, processes, and quantities are more intense, complex, or abundant in ponds and small lakes than in larger lakes. The biotic complexity and richness of small aquatic systems is well-known. For example, macrophyte coverage is greater in smaller lakes (Duarte *et al.* 1986) leading to enhanced production and habitat composition. In the pelagic zone, too, small lakes have more complex thermal structure than large ones (Xenopoulos & Schindler 2001).

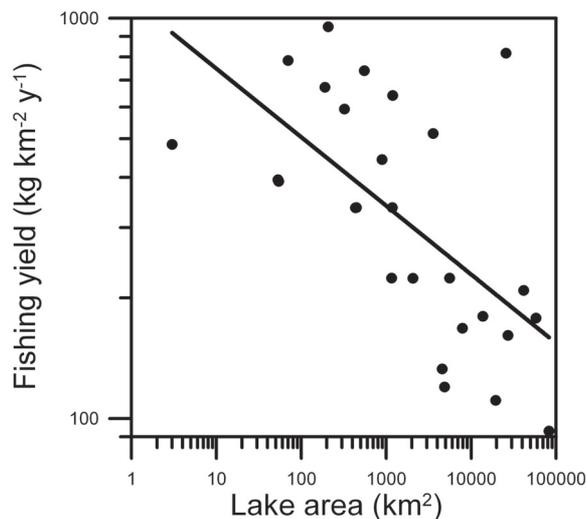
Small lakes and ponds are important to the maintenance of regional biodiversity and stability. Small lakes have greater waterfowl species richness per unit area than large lakes (Elmberg *et al.* 1994). Small lakes and ponds promote enhanced regional biodiversity in aquatic birds, plants, amphibians and invertebrates because of low fish biomass and high richness and abundance of aquatic plants (Scheffer *et al.* 2006). Smaller lakes have a greater proportion of small non-game fish species such as the Cyprinidae (Matuszek *et al.* 1990); small non-game fish are often overlooked by fish management. Biomass size spectra show more negative coefficients in small lakes indicating a greater dominance of small, active organisms (Cyr & Peters 1996). Figure 5 shows data on biodiversity in well-studied lakes analyzed by Dodson *et al.* (2000). The data indicate that small lakes contain many more species of virtually all taxa, per unit area, than do large lakes. Although no particular meaning should be attributed to the existence of such a correlation ( $\text{km}^2$  appears in both axes), even moderate



**Figure 5.** Species-richness per unit area of various aquatic taxa in lakes of different sizes (Data from Dodson *et al.* 2000). If individual lakes in the same region have slightly different community structure, the figure implies that small systems enhance regional biodiversity. *Riqueza de especies por unidad de superficie de varios grupos taxonómicos en lagos de diferente tamaño (Datos de Dodson et al. 2000). Si los lagos individuales de una región tienen comunidades ligeramente diferentes, la figura indica que los sistemas pequeños aumentan la biodiversidad regional.*

differences in community structure among small lakes and ponds suggest that higher regional biodiversity can be maintained by 100 km<sup>2</sup> of small lakes than would be contributed by a single 100 km<sup>2</sup> lake. This, plus the preference of recreational boaters for large lakes (Reed Andersen *et al.* 2000), may help explain why small lakes are known to be more resistant to invasion by exotic and nuisance species than are large ones (Winfield *et al.* 1998).

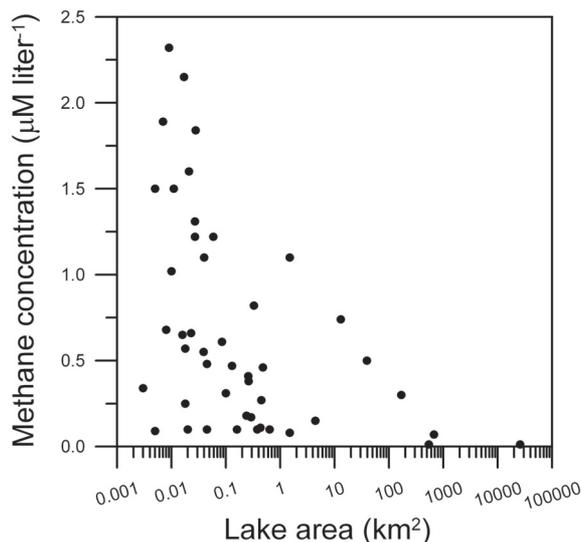
Small lakes and ponds are also known for high productivity. Fish productivity generally declines with increasing lake size, indicating that smallest lakes have highest production per unit area, often by several orders of magnitude (Rounsefell 1946, Hayes & Anthony 1964, Youngs & Heimbuch 1982, Downing *et al.* 1990) (Fig. 6). Lake size appears to act on biomass and fish-size distribution because after the effects of body mass and biomass are accounted for, fish production (per unit area) may be higher in larger lakes (Downing & Plante 1993). Small lakes and ponds can be substantially more biologically active than large lakes.



**Figure 6.** Fish yield and lake-size data summarized by Youngs & Heimbuch (1982) from other sources (Ryder 1965, Oglesby 1977, Matuszek 1978). The solid line is a least-squares regression of the data showing the average trend in production with lake size ( $r^2 = 0.39$ ,  $n = 27$ ). *Producción pesquera en relación con el tamaño del lago. Datos recogidos por Youngs & Heimbuch (1982) de diversas fuentes (Ryder 1965, Oglesby 1977, Matuszek 1978). La línea sólida representa la regresión por mínimos cuadrados, mostrando la relación de la producción con el tamaño del lago ( $r^2 = 0.39$ ,  $n = 27$ ).*

### Carbon-processing is intense in small lakes and ponds

Information is beginning to emerge showing that carbon processing intensity is very great in small water bodies. Stable isotope analyses indicate that smaller lakes and ponds may be more heterotrophic than large ones, processing substantial amounts of terrestrial or external carbon (Post 2002). Dissolved organic carbon concentrations are therefore significantly negatively correlated with lake size (Xenopoulos *et al.* 2003). Surface CO<sub>2</sub> concentrations are much higher in smaller lakes than large ones (Kelly *et al.* 2001). In another large data set taken from across Finland, CO<sub>2</sub> concentrations and aerial CO<sub>2</sub> evasion declined sharply with increasing lake size (Kortelainen *et al.* 2006). Oxygen concentrations tend to be lower in ponds and small lakes than in larger ones (Crisman *et al.* 1998), enhancing greenhouse gas (GHG) emissions and carbon sequestration. Potential methane emission is much



**Figure 7.** Measured methane concentrations in lakes from around the world related to the sizes of lakes. Data are from Bastviken *et al.* (2004). *Concentraciones de metano en lagos de diferentes partes del mundo, en relación con el tamaño de los lagos.* (Datos de Bastviken *et al.* 2004).

greater in small lakes than large ones (Michmerhuizen *et al.* 1996). Using a data compilation from around the world, Bastviken *et al.* (2004) showed that concentrations of methane, and perhaps therefore losses to the atmosphere, are greatest in small lakes and ponds (Fig. 7). Low oxygen concentrations in small lakes (Crisman *et al.* 1998) and the relationship between low oxygen and elevated  $N_2O$  (Knowles *et al.* 1981) suggest that  $N_2O$  emissions from ponds and small lakes can be much higher than those of larger lakes. Rates of organic carbon sequestration per unit area in the sediments of small lakes has been suggested to be at least an order of magnitude higher than that of larger lakes (Dean & Gorham 1998, Stallard 1998, Downing *et al.* 2008).

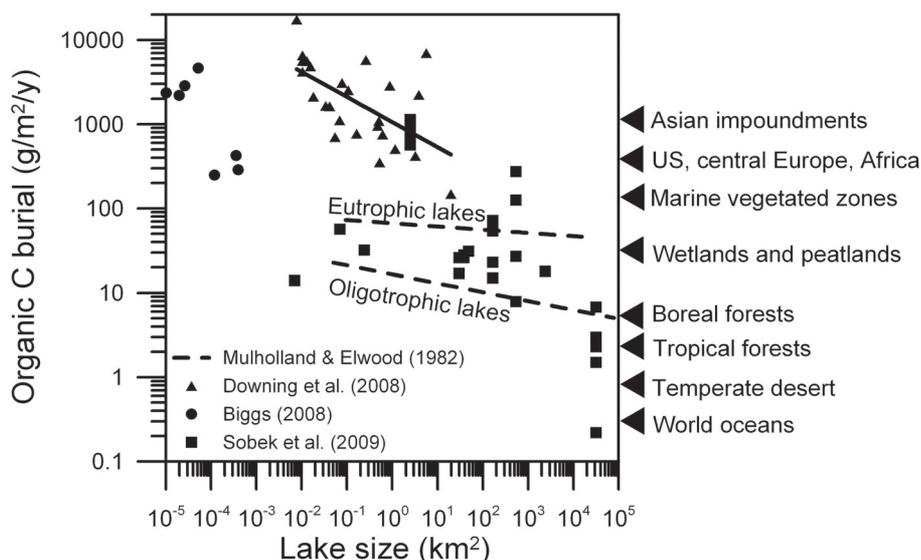
### **Pond size, eutrophication, and carbon sequestration: some examples**

The global importance of an aquatic process or quantity depends, to some degree, upon the extent of the ecosystem type in the biosphere. Likewise, seemingly unimportant ecosystems, even those that cover only a small area of the land sur-

face, can be important globally if the intensity of a process is extremely high. Even the smallest ponds are very abundant on Earth. A conservative estimate is that small agricultural ponds cover about 77,000 km<sup>2</sup> worldwide (Downing *et al.* 2006, Downing & Duarte 2009). Farm ponds and tanks appear to be increasing at rates from 0.7 % per year to 60 % per year in various regions as increasing pressure is put on agricultural lands to provide food for growing populations.

Previous analyses of roles of constructed lakes in important global rates like organic C burial (e.g., Cole *et al.* 2007) have calculated global deposition and carbon content of sediments derived mostly from large water bodies (Dendy & Champion 1978, Mulholland & Elwood 1982, Dean & Gorham 1998, Stallard 1998). Because these data seemed limited and ignored the active and abundant small lakes and ponds on Earth, we recently used repeated bathymetric analyses and direct measures of sediment characteristics to estimate the likely rate of burial of organic C in the sediments of eutrophic lakes and impoundments (Downing *et al.* 2008). In the 40 lakes we studied (triangles, Fig. 8), we found that sediment organic carbon burial rates were higher than those assumed for fertile impoundments by previous studies and were much higher than those measured in natural lakes. Organic carbon burial ranged from a high of 17 kg C/m<sup>2</sup>/y to a low of 148 g C/m<sup>2</sup>/y and was significantly greater in small impoundments than large ones (Fig. 8).

These analyses suggest that median organic C sequestration in moderate to large impoundments may be double the rate assumed in previous analyses and exceeds rates of carbon sequestration found in any ecosystem in the world. Median areal C burial rates in these lakes were 10-times those seen in wetlands, 100-times those documented in tropical forests, 1000-times those assessed in tropical and boreal forests, and 10,000-times those estimated for the world's oceans. Extrapolation suggests that each year, Earth's current moderately sized impoundments may bury 4-times as much C as the world's oceans. The world's farm ponds alone seem likely to sequester more organic carbon each year than the oceans and 33 % as much as the world's rivers deliver to the sea.



**Figure 8.** Sediment organic carbon burial rates compared among types of aquatic and terrestrial ecosystems. Data on oligotrophic and eutrophic lakes and impoundments in Asia, the United States, central Europe, and Africa are from Mullholland & Elwood (1982). Data from Downing *et al.* (2008) are for lakes in an agriculturally eutrophic region of the Midwest United States; the solid line shows a least squares regression of these data. Observations made by Biggs (2008) are for small ponds in the United Kingdom. Data from Sobek *et al.* (2009) include a variety of lakes worldwide, including Lake Baikal at the extreme right of the graph. Terrestrial data are from Schlesinger (1997) and data on marine vegetated areas are from Duarte *et al.* (2005). Carbon burial in the world's oceans were calculated after Sarmiento & Sundquist (1992) assuming the world's oceans have an area of 361 million km<sup>2</sup>. Arrows at right indicate median levels of carbon sequestration in diverse ecosystem types. *Comparación de las tasas de entierro de carbono en diferentes tipos de ecosistemas acuáticos y terrestres. Los datos de lagos y embalses oligotróficos y eutróficos de Asia, Estados Unidos, Europa Central y Africa proceden de Mullholland & Elwood (1982). Los datos de Downing et al. (2008) corresponden a lagos en una región agrícola y eutrófica del Oeste Medio de Estados Unidos y la línea sólida representa la regresión por mínimos cuadrados de estos datos. Las observaciones de Biggs (2008) corresponden a pequeñas charcas del Reino Unido. Los datos de Sobek et al. (2009) incluyen una variedad de lagos de todo el mundo, con el lago Baikal en el extremo derecho del gráfico. Los datos terrestres son de Schlesinger (1997) y los de áreas marinas vegetadas de Duarte et al. (2005). El entierro de carbono en los océanos se ha calculado de acuerdo con Sarmiento & Sundquist (1992) asumiendo que los océanos ocupan una superficie 361 millones de km<sup>2</sup>. Las flechas de la derecha indican la mediana de los niveles de secuestro de carbono en diversos tipos de ecosistemas.*

Eutrophication and landscape alteration may play important roles in determining C burial in lakes. C burial rates in eutrophic lakes are nearly an order of magnitude higher than those found in oligotrophic lakes of similar size (Fig. 8). Small lakes in agricultural regions (Downing *et al.* 2008) have very high rates of burial but are in the same range as the small UK ponds, impoundments around the world, and lakes with high sediment loads. For example, Lake Wohlen (Sobek *et al.* 2009), a mesotrophic, short water residence time (2 days) impoundment in the Aare River has C sequestration rates of 570-1140 g C/m<sup>2</sup>/y. Therefore, it appears that extremely high rates of C burial are typical of small lakes, lakes with high rates of primary production due to eutroph-

ication, and lakes receiving substantial loads of riverine or watershed-derived organic sediments. Small lakes and ponds make up around a third of the area of continental waters but have rates of C burial that exceed those of larger lakes by an order of magnitude or more. It is likely, therefore, that carbon sequestration by the world's small lakes and ponds dominates carbon burial by aquatic ecosystems. Because aquatic ecosystems seem to provide substantial carbon burial worldwide, ponds and small lakes may be the most important sites in the biosphere for organic carbon sequestration.

These findings should not be misconstrued to suggest that small lakes and ponds are perfect sinks for excess carbon. Small oligotrophic lakes may evade substantial allochthonous C as CO<sub>2</sub>

(Kelly *et al.* 2001, Kortelainen *et al.* 2006). Small lakes and ponds can be quite eutrophic so CH<sub>4</sub> and N<sub>2</sub>O release may be substantial (Knowles *et al.* 1981, Michmerhuizen *et al.* 1996, Bastviken *et al.* 2004), exacerbating atmospheric problems. This analysis suggests, however, that an accurate view of the global carbon budget will be elusive unless small lakes and ponds are analyzed, understood, and considered.

### Global research needs for small aquatic ecosystems

Global understanding of the role of small lakes and ponds in processes throughout the biosphere requires inventories of water bodies and knowledge of the important rates and processes they mediate (Downing 2009). There are three important steps. (1) We need to identify patterns in globally important quantities, rates, and processes, and understand how they covary with lake and pond characteristics. (2) We need to create scaling rules for these quantities, rates, and processes that will permit meaningful up-scaling to a global level. (3) Because society depends upon reliable global science, we need to derive numerical and statistical methods to ensure that global calculations are accurate and precise enough to be comparable to other global estimates. Accomplishment of these tasks will advance us substantially toward estimating human- and climate-mediated effects on the global role of small aquatic ecosystems.

Many variables are in need of global scaling. For example, understanding the conversions of carbon in small lakes and ponds is of very high priority, in order to contribute substantially to discussions of global climate change. Likewise, understanding of patterns in nutrients in these water bodies, as well as fluxes and conversions of important gasses (e.g., N<sub>2</sub>O, NH<sub>x</sub>) and metals (e.g., Hg), will improve global understanding of the role of small water bodies in global nutrient, gas, and toxin budgets. Remarkably, small lakes and ponds have not yet been integrated into global heat and water budgets so recognition of patterns in water and energy fluxes amongst aquatic systems is also important. Small aquatic ecosystems are disproportionately important sites

for the production of food so it is important to evaluate global patterns in production.

We need to quantify and understand the role of small water bodies in the functioning of the biosphere. We do this by asking whether the quantity or process is large or small with respect to other types of ecosystems and whether we can make an estimate of that quantity or process that is well enough constrained to be reliable. These questions cause us to ascertain whether the process is likely great enough to justify a more accurate and precise answer and how likely we are to be able to define the answer more precisely. Therefore, much of this task is making estimates of biosphere-level rates and processes attributable to small lakes and ponds, comparing these to estimates made for other ecosystems, and refining and improving our estimates to yield more accurate and precise assessments of the global role of small aquatic systems.

### CONCLUSIONS

Recently, limnologists and aquatic ecologists have discovered that aquatic ecosystems are much more plentiful in the biosphere than had been believed. This is especially true for small lakes and ponds because new analyses show that they cover as much or more area as large lakes. Because historical inventories underestimated the areal extent of small water bodies, limnologists have spent relatively little effort studying them so their importance to global and biosphere processes has been under-appreciated. Emerging studies now show that ponds and small lakes are more active in nearly every process than large lakes, terrestrial, and marine ecosystems. The large area covered by small aquatic systems and the intensity of activity mean that they may be among the most important ecosystems in the world. Considering the global carbon cycle, for example, ponds and small lakes sequester carbon at rates that are orders-of-magnitude greater than virtually all other global ecosystems. This compensates for the small area they cover relative to terrestrial and marine ecosystems, suggesting that carbon sequestration by ponds may

be as great as or greater than that of forests, grasslands, and all the world's oceans. There are several knowledge gaps, however, including information on gas evasion and several other factors, so an active research agenda on small lakes and ponds is needed to bring them into the arena of global limnology and ecology. Work in such a high-priority arena is important to our science and careers but especially to understanding the role of small aquatic systems in the biosphere. Preliminary information suggests that they may be amongst Earth's most important and active environments.

## ACKNOWLEDGEMENTS

I am grateful to the European Pond Conservation Network 2008 organizing committee for inviting me to address this important topic. I am also grateful to the NCEAS-ITAC group (authors of Downing *et al.* 2006, Cole *et al.* 2007), for advancing many of the subjects presented here. This work grew out of the ITAC Working Group supported by the National Center for Ecological Analysis and Synthesis, a Center funded by NSF (Grant DEB-94-21535), the University of California at Santa Barbara, and the State of California. This work was partially completed while I was on a sabbatical leave at Instituto Mediterraneo de Estudios Avanzados, Esporles, Mallorca, Islas Baleares, Spain, with the generous sponsorship of the Consejo Superior de Investigaciones Científicas of Spain. Other support was provided by the Wabana Lake Research Station.

## REFERENCES

- AMIEL, H.-F. 1893. *Journal Intime*. Macmillan. London, U.K. 402 pp.
- BARTON, B. F. 1917. *More power to you: fifty editorials from every week*. The Century Company. New York, USA. 232 pp.
- BASTVIKEN, D., J. J. COLE, M. L. PACE & L. J. TRANVIK 2004. Methane emissions from lakes: Dependence of lake characteristics, two regional assessments, and a global estimate. *Global Biogeochemical Cycles* 18 10.1029/2004GB002238: 12.
- BIGGS, J. 2008. Carbon uptake by UK ponds. Pond Conservation for Life in Freshwater. May 6, 2008. Oxford, UK. [<http://www.pondconservation.org.uk/aboutus/ourwork/climatechangeandponds/carbonuptakebyUKponds>, accessed in September, 2009]
- BIN-LE, L., A. SAKODA, R. SHIBASAKI, N. GOTO & M. SUZUKI. 2000. Modelling a global biogeochemical nitrogen cycle in terrestrial ecosystems for the evaluation of anthropogenic disturbance. *Ecological Modelling*, 135: 89-110.
- BOLIN, B. 1983. The carbon cycle. In: *The major biogeochemical cycles and their interactions*, B. Bolin & R. B. Cook (eds.): 980-987. John Wiley & Sons, New York, USA.
- CHAMEIDES, W. L. & E. M. PERDUE. 1997. *Biogeochemical Cycles: A Computer-Interactive Study of Earth System Science and Global Change*. Oxford University Press. New York, USA. 240 pp.
- CHRISTOPHERSON, R. W. 1994. *Geosystems: An Introduction to Physical Geography*, 2nd ed. Prentice Hall. Upper Saddle River, New Jersey, USA. 798 pp.
- CLARKE, R. 1991. *Water: the international crisis*. Earthscan Publications Limited. London, UK. 224 pp.
- CLOUD, P. & A. GIBOR. 1970. The oxygen cycle. *Scientific American*, September 110-123.
- COLE, J. J., Y. T. PRAIRIE, N. F. CARACO, W. H. MCDOWELL, L. J. TRANVIK, R. G. STRIEGL, C. M. DUARTE, P. KORTELAINEN, J. A. DOWNING, J. MIDDELBURG & J. M. MELACK. 2007. Plumbing the global carbon cycle: integrating inland waters into the terrestrial carbon cycle. *Ecosystems*, 10: 171-184.
- CONAN DOYLE, A. 1920. *Adventures of Sherlock Holmes*. A. L. Burt Company. New York, NY, USA. 307 pp.
- CRISMAN, T. L., L. J. CHAPMAN & C. A. CHAPMAN. 1998. Predictors of seasonal oxygen levels in small Florida lakes The importance of color. *Hydrobiologia*, 368: 149-155.
- CYR, H. & R. H. PETERS. 1996. Biomass size spectra and the prediction of fish biomass in lakes. *Canadian Journal of Fisheries and Aquatic Sciences*, 53: 685-697.
- DEAN, W. E. & E. GORHAM. 1998. Magnitude and significance of carbon burial in lakes, reservoirs, and peatlands. *Geology*, 26: 535-538.
- DENDY, F. E. & W. A. CHAMPION. 1978. Sediment deposition in U.S. reservoirs: summary of

- data reported through 1975. In: *United States Department of Agriculture, Miscellaneous Publication: 84*. Agricultural Research Service, United States Department of Agriculture, Washington, D.C.
- DODSON, S. I., S. E. ARNOTT & K. L. COTTINGHAM. 2000. The relationship in lake communities between primary productivity and species richness. *Ecology*, 81: 2662-2679.
- DOWNING, J. A. 2009. Global limnology: Up-scaling aquatic services and processes to planet Earth. *Verh. Internat. Verein. Limnol.*, 30: 1149-1166.
- DOWNING, J. A., J. J. COLE, J. MIDDELBURG, R. G. STRIEGL, C. M. DUARTE, P. KORTELAINEN, Y. T. PRAIRIE & K. A. LAUBE. 2008. Sediment carbon burial in agriculturally eutrophic impoundments over the last century. *Global Biogeochemical Cycles*, 22: 10.1029/2006GB002854.
- DOWNING, J. A. & C. M. DUARTE. 2009. Abundance and size distribution of lakes, ponds, and impoundments. In: *Encyclopedia of Inland Waters*, G. E. Likens (ed.): 469-478. Elsevier, Oxford, U.K.
- DOWNING, J. A. & C. PLANTE. 1993. Production of fish populations in lakes. *Canadian Journal of Fisheries and Aquatic Sciences*, 50: 110-120.
- DOWNING, J. A., C. PLANTE & S. LALONDE. 1990. Fish production correlated with primary productivity, not the morphoedaphic index. *Canadian Journal of Fisheries and Aquatic Sciences*, 47: 1929-1936.
- DOWNING, J. A., Y. T. PRAIRIE, J. J. COLE, C. M. DUARTE, L. J. TRANVIK, R. G. STRIEGL, W. H. MCDOWELL, P. KORTELAINEN, N. F. CARACO, J. M. MELACK & J. MIDDELBURG. 2006. The global abundance and size distribution of lakes, ponds, and impoundments. *Limnology and Oceanography*, 51: 2388-2397.
- DUARTE, C. M., J. KALFF & R. H. PETERS. 1986. Patterns in biomass and cover of aquatic macrophytes in lakes. *Canadian Journal of Fisheries and Aquatic Sciences*, 43: 1900-1908.
- DUARTE, C. M., J. J. MIDDELBURG & N. F. CARACO. 2005. Major role of marine vegetation on the oceanic carbon cycle. *Biogeosciences*, 2: 1-8.
- ELMBERG, J., P. NUMMI, H. POYSA & K. SJOBERG. 1994. Relationships between species number, lake size and resource diversity in assemblages of breeding waterfowl. *Journal of Biogeography*, 21: 75-84.
- FRENEY, J. R., M. V. IVANOV & H. RODHE. 1983. The sulphur cycle. In: *The major biogeochemical cycles and their interactions*, B. Bolin & R. B. Cook (eds.): 56-61. John Wiley & Sons, New York, USA.
- FRIEDERICI, H. (ed.). 1978. *Lichtenbergs Werke in einem Band*. Aufbau-Verlag, Berlin, Germany. 378 pp.
- GOODY, R. M. & J. C. G. WALKER. 1972. *Atmospheres*. Prentice-Hall. Englewood Cliffs, NJ, USA. 160 pp.
- GORHAM, E. 1958. The physical limnology of Great Britain: an epitome of the bathymetric survey of the Scottish Freshwater Lochs. *Limnology and Oceanography*, 3: 40-50.
- GRAHAM, W. F. & R. A. DUCE. 1979. Atmospheric pathways of the phosphorus cycle. *Geochimica Cosmochimica Acta*, 43: 1195-1208.
- HALBFASS, W. 1914. *Das Süßwasser der Erde (The freshwater of the Earth)*. Druck und Verlag von Philipp Reclam jun. Leipzig, Germany. 189 pp.
- HAYES, F. R. & E. H. ANTHONY. 1964. Productive capacity of North American lakes as related to the quantity and the trophic level of fish, the lake dimensions, and the water chemistry. *Transactions of the American Fisheries Society*, 93: 53-57.
- HERDENDORF, C. E. 1984. Inventory of the morphometric and limnologic characteristics of the large lakes of the world. In: *Technical Bulletin*, 1-54. The Ohio State University Sea Grant Program.
- HERMANN, W. 2006. Quantifying global exergy resources. *Energy*, 31: 1349-1366.
- HINRICHSSEN, D., B. ROBEY & U. D. UPADHYAY. 1998. *Solutions for a water-short world*. Population Information Program, Center for Communication Programs, The Johns Hopkins School of Public Health. Baltimore, USA. 31 pp.
- INTERGOVERNMENTAL PANEL ON CLIMATE CHANGE. 2001. *Climate change 2001: the scientific basis*. Cambridge University Press. Cambridge, UK. 881 pp.
- KALFF, J. 2001. *Limnology: inland water ecosystems*. Prentice Hall. Upper Saddle River, NJ, USA. 592 pp.
- KEELING, R. F., R. P. NAJJAR, M. L. BENDER & P. P. TANS. 1993. What atmospheric oxygen measurements can tell us about the global carbon cycle. *Global Biogeochemical Cycles*, 7: 37-67.
- KELLY, C. A., E. J. FEE, P. S. RAMLAL, J. W. M. RUDD, R. H. HESSLEIN, C. ANEMA & E. U. SCHINDLER. 2001. Natural variability of carbon

- dioxide and net epilimnetic production in the surface waters of boreal lakes of different sizes. *Limnology and Oceanography*, 46: 1054-1064.
- KIEHL, J. T. & K. E. TRENBERTH. 1997. Earth's annual global mean energy budget. *Bulletin of the American Meteorological Association*, 78: 197-208.
- KNOWLES, R., D. R. S. LEAN & Y. K. CHAN. 1981. Nitrous oxide concentrations in lakes: variations with depth and time. *Limnology and Oceanography*, 26: 855-866.
- KORTELAJINEN, P., M. RANTAKARI, J. T. HUTTUNEN, T. MATTSSON, J. ALM, S. JUUTINEN, T. LARMOLA, J. SILVOLA & P. J. MARTIKAINEN. 2006. Sediment respiration and lake trophic state are important predictors of large CO<sub>2</sub> evasion from small boreal lakes. *Global Change Biology*, 12 10.1111/j.1365-2486.2006.01167.x: 1554-1567.
- LEHNER, B. & P. DÖLL. 2004. Development and validation of a global database of lakes, reservoirs and wetlands. *Journal of Hydrology*, 296: 1-22.
- LERMAN, A. 1988. *Geochemical Processes-Water and Sediment Environments*. Krieger Publishing Company. Malabar, Florida, USA. 492 pp.
- MATUSZEK, J. E. 1978. Empirical predictions of fish yields of large North American lakes. *Transactions of the American Fisheries Society*, 107: 385-394.
- MATUSZEK, J. E., J. GOODIER & D. L. WALES. 1990. The occurrence of Cyprinidae and other small fish species in relation to pH in Ontario lakes. *Transactions of the American Fisheries Society*, 119: 850-861.
- MEYBECK, M. 1995. Global distribution of lakes. In: *Physics and chemistry of lakes*, A. Lerman, D. M. Imboden & J. R. Gat (eds.): 1-35. Springer-Verlag, Berlin, Germany.
- MICHMERHUIZEN, C. M., R. G. STRIEGL & M. E. MCDONALD. 1996. Potential methane emission from north-temperate lakes following ice melt. *Limnology and Oceanography*, 41: 985-991.
- MULHOLLAND, P. J. & J. W. ELWOOD. 1982. The role of lake and reservoir sediments as sinks in the perturbed global carbon cycle. *Tellus*, 34: 490-499.
- NELSON, D. M., P. TRÉGUER, M. A. BRZEZINSKI, A. LEYNAERT & B. QUÉGUINER. 1995. Production and dissolution of biogenic silica in the ocean: Revised global estimates, comparison with regional data and relationship to biogenic sedimentation. *Global Biogeochemical Cycles*, 9: 359-372.
- OERTLI, B., R. CÉRÉGHINO, A. HULL & R. MIRACLE. 2009. Pond conservation: from science to practice. *Hydrobiologia*, 10.1007/s10750-009-9891-9.
- OGLESBY, R. T. 1977. Relationship of fish yield to lake phytoplankton standing crop, production, and morphoedaphic factors. *Journal of the Fisheries Research Board of Canada*, 34: 2271-2279.
- PARETO, V. 1897. *Cours d'économie politique*. F. Rouge. 2 v. Lausanne, Switzerland. 426 pp.
- POST, D. M. 2002. Using stable isotopes to estimate trophic position: models, methods, and assumptions. *Ecology*, 83: 703-718.
- RAVEN, P. H., G. B. JOHNSON, J. B. LOSOS & R. SINGER. 2004. *Biology*, Sixth ed. McGraw-Hill Higher Education. Boston, USA. 1238 pp.
- REED ANDERSEN, T., E. M. BENNETT, B. S. JORGENSEN, G. LAUSTER, D. B. LEWIS, D. NOWACEK, J. L. RIERA, B. L. SANDERSON & R. STEDMAN. 2000. Distribution of recreational boating across lakes: Do landscape variables affect recreational use? *Freshwater Biology*, 43: 439-448.
- RICHEY, J. E. 1983. The phosphorus cycle. In: *The major biogeochemical cycles and their interactions*, B. Bolin & R. B. Cook (eds.): 51-56. John Wiley & Sons. New York, USA.
- ROSSWALL, T. 1983. The nitrogen cycle. In: *The major biogeochemical cycles and their interactions*, B. Bolin & R. B. Cook (eds.): 46-50. John Wiley & Sons. New York, USA.
- ROUNSEFELL, G. A. 1946. Fish production in lakes as a guide for estimating production in proposed reservoirs. *Copeia*, 1: 29-40.
- ROY, R. N., R. V. MISRA, J. P. LESSCHEN & E. M. SMALING. 2003. *Assessment of soil nutrient balance: approaches and methodologies*. Food and Agriculture Organization of the United Nations. Rome, Italy. 98 pp.
- RYDER, A. 1965. A method for estimating the potential fish production of north-temperate lakes. *Transactions of the American Fisheries Society*, 94: 214-218.
- SARMIENTO, J. L. & E. T. SUNDQUIST. 1992. Revised budget for the oceanic uptake of anthropogenic carbon dioxide. *Nature*, 356: 589-593.
- SCHEFFER, M., G. J. VAN GEEST, K. ZIMMER, M. G. BUTLER, M. A. HANSON, S. DECLERCK, L. DE MEESTER, E. JEPPESEN & M. SONDERGAARD. 2006. Small habitat size and

- isolation can promote species richness: second-order effects on biodiversity in shallow lakes and ponds. *Oikos*, 112: 227-231.
- SCHIMMEL, D. S., I. ENTING, M. HEIMANN, T. M. WIGLEY, D. RAYNAUD, D. ALVES, & U. SIEGENTHALER. 1995. CO<sub>2</sub> and the carbon cycle. In: *Climate Change : Radiative Forcing of Climate Change and An Evaluation of the IPCC IS92 Emission Scenarios*, J. T. Houghton (ed.): 35-71. Cambridge University Press, Cambridge, UK.
- SCHLESINGER, W. H. 1997. *Biogeochemistry: an analysis of global change*. Academic Press. San Diego, California, USA. 588 pp.
- SCHUILING, R. D. 1977. Source and composition of lake sediments. In: *Interaction between sediments and fresh water: Proceedings of an international symposium held at Amsterdam, the Netherlands, September 6-10, 1976*, H. L. Golterman (ed.): 12-18. Dr. W. Junk B.V., Wageningen, Netherlands.
- SEINFELD, J. H. & S. N. PANDIS. 1998. *Atmospheric chemistry and physics: from air pollution to climate change*. Wiley and Sons. New York, USA. 1326 pp.
- SHIKLOMANOV, I. A. & J. C. RODDA. 2003. *World water resources at the beginning of the twenty-first century*. Cambridge University Press. Cambridge, U.K. 452 pp.
- SOBEK, S., E. DURISCH-KAISER, R. ZURBRÜGG, N. WONGFUN, M. WESSELS, N. PASCHE & B. WEHRLI. 2009. Organic carbon burial efficiency in lake sediments controlled by oxygen exposure time and sediment source. *Limnology and Oceanography*, 54: 2243-2254.
- STALLARD, R. F. 1998. Terrestrial sedimentation and the C cycle: coupling weathering and erosion to carbon storage. *Global Biogeochemical Cycles*, 12: 231-237.
- STRAŠKRABA, M. 1980. The effects of physical variables on freshwater production: analyses based on models. In: *The functioning of freshwater ecosystems*, E. D. Le Cren & R. H. Lowe-McConnell (eds.): 13-84. International Biological Programme. Cambridge University Press, Cambridge, U.K.
- THIENEMANN, A. 1925. *Die Binnengewässer Mitteleuropas (The inland water of Central Europe)*. E. Schweizerbart'sche Verlagsbuchhandlung. Stuttgart, Germany. 255 pp.
- THORNELOE, S. A., K. A. WEITZ, S. R. NISHITALA, S. YARKOSKY & M. ZANNES. 2002. The impact of municipal solid waste management on greenhouse gas emissions in the United States. *Journal of Air and Waste Management Association*, 52: 1000-1011.
- TRANVIK, L. J., J. A. DOWNING, J. B. COTNER, S. A. LOISELLE, R. G. STRIEGL, T. J. BALLATORE, P. J. DILLON, K. FINLAY, K. FORTINO, L. B. KNOLL, P. KORTELAINEN, T. KUTSER, S. LARSEN, I. LAURION, D. M. LEECH, S. L. MCCALLISTER, D. M. MCKNIGHT, J. M. MELACK, E. OVERHOLT, J. A. PORTER, Y. T. PRAIRIE, W. H. RENWICK, F. ROLAND, B. S. SHERMAN, D. W. SCHINDLER, S. SOBEK, A. TREMBLAY, M. J. VANNI, A. M. VERSCHOOR, E. VON WACHENFELDT & G. A. WEYENMEYER. 2009. Lakes and impoundments as regulators of carbon cycling and climate. *Limnology and Oceanography*, 54: 2298-2314.
- TRÉGUER, P., D. M. NELSON, A. J. V. BENEKOM, D. J. DEMASTER, A. LEYNAERT & B. QUÉGUINER. 1995. The silica balance in the world ocean: a reestimate. *Science*, 268: 375-379.
- UNITED STATES CLIMATE CHANGE SCIENCE PROGRAM. 2003. *Strategic plan for the U.S. climate change science program*. United States Government Printing Office. Washington, DC. 364 pp.
- VIDONDO, B., Y. T. PRAIRIE, J. M. BLANCO & C. M. DUARTE. 1997. Some aspects of the analysis of size spectra in aquatic ecology. *Limnology and Oceanography*, 42: 184-192.
- WALKER, J. C. G. 1980. The Oxygen Cycle. In: *The Natural Environment and the Biogeochemical Cycles*. O. Hutzinger (ed.): 87-104. Springer-Verlag. Berlin, Germany.
- WEISSERT, H. 2000. Global change: deciphering methane's fingerprint. *Nature*, 406: 356-357.
- WETZEL, R. W. 1990. Land-water interfaces: metabolic and limnological regulators. *Verh. Internat. Verein. Limnol.*, 24: 6-24.
- WETZEL, R. W. 2001. *Limnology: lake and river ecosystems*. Academic Press. San Diego, CA, USA. 1006 pp.
- WINFIELD, I. J., R. ROESCH, M. APPELBERG, A. KINNERBACK & M. RASK. 1998. Recent introductions of the ruff (*Gymnocephalus cernuus*) to *Coregonus* and *Perca* Lakes in Europe and an analysis of their natural distributions in Sweden and Finland. *Journal of Great Lakes Research*, 24: 235-248.
- WINTER, T. C., J. W. HARVEY, O. L. FRANKE & W. M. ALLEY. 1998. *Ground water and surface*

- water, a single resource*. United States Geological Survey. 79 pp.
- XENOPOULOS, M. A., D. M. LODGE, J. FREN-  
TRESS, T. A. KREPS, S. D. BRIDGHAM, E.  
GROSSMAN & C. J. JACKSON. 2003. Re-  
gional comparisons of watershed determinants  
of dissolved organic carbon in temperate lakes  
from the Upper Great Lakes Region and selected  
regions globally. *Limnology and Oceanography*,  
48: 2321-2334.
- XENOPOULOS, M. A. & D. W. SCHINDLER.  
2001. The environmental control of near surface  
thermoclines in boreal lakes. *Ecosystems*, 4: 699-  
707.
- YOUNGS, W. D. & D. G. HEIMBUCH. 1982. An-  
other consideration of the morphoedaphic index.  
*Transactions of the American Fisheries Society*,  
111: 151-153.